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ORIGINAL ARTICLE

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Silica-encapsulated DNA tracers for measuring aerosol distribution dynamics in real-world settings

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Abstract

Aerosolized particles play a significant role in human health and environmental risk management. The global importance of aerosol-related hazards, such as the circulation of pathogens and high levels of air pollutants, have led to a surging demand for suitable surrogate tracers to investigate the complex dynamics of airborne particles in real-world scenarios. In this study, we propose a novel approach using silica particles with encapsulated DNA (SPED) as a tracing agent for measuring aerosol distribution indoors. In a series of experiments with a portable setup, SPED were successfully aerosolized, recaptured, and quantified using quantitative polymerase chain reaction (qPCR). Position dependency and ventilation effects within a confined space could be shown in a quantitative fashion achieving detection limits below 0.1 ng particles per m³ of sampled air. In conclusion, SPED show promise for a flexible, cost-effective, and low-impact characterization of aerosol dynamics in a wide range of settings.

KEYWORDS

aerosol, DNA, nanoparticles, silica, tracing, ventilation

1 | INTRODUCTION

Aerosol dynamics are an important factor when assessing the circulation of hazardous pollutants and pathogens with regard to human health and the environment. Viable bioaerosols are a well-known cause for many infectious diseases such as tuberculosis,¹ measles,² Legionnaire's disease,³ influenza,⁴ gastroenteritis,⁵ and SARS-CoV-1⁶ and SARS-CoV-2.^{7–10} With the COVID-19 pandemic spreading globally, the importance of understanding the mechanisms of aerosol spreading has perhaps become more evident than ever, and discussions surrounding the topic have reached the mainstream. Specifically, the assessment and characterization of aerosol distribution properties within specific indoor spaces such as lecture halls, office spaces, hospitals, public transport vehicles, or event venues are of high interest with regard to public and occupational health.

Many of the existing approaches to assess indoor environments regarding aerosol dynamics rely on computational models. However, such models are based on the preexisting understanding of fluid dynamics and prone to numerical error,¹¹ while also depending on data for setting realistic parameters.¹² Therefore, predicting real-world behavior in complex, non-controlled environments is computationally challenging. For those reasons, physical tracing is still a vital field to assess aerosol-related properties in real-world settings to help validate and complement model-based studies.

Table 1 shows a list of studies reporting tracing methods to characterize (bio-)aerosol distribution indoors. Carbon dioxide, for example, is a commonly used tracer gas, applied either alone or in combination with other tracers, and in conjunction with one or more CO₂ detectors. For example, Knibbs et al.¹³ employed CO₂ to assess the effect of ventilation rates in various room settings, using the results to model virus-specific infection risks. While CO₂ is widely accessible,

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non-toxic (in low concentrations), easily detected, and self-clearing, it is also naturally present in ambient air. Subtle effects may therefore be harder to detect, and relatively high working concentrations are needed. Moreover, aerosol dynamics are complex and not necessarily approximated accurately by a non-particulate tracer gas.

Schade et al.¹⁴ therefore combined CO₂ with an aerosol tracer to characterize distribution in a large concert hall. Their combinatorial approach has the advantage of monitoring aerosol droplets with temporal and spatial resolution, with the limitation that the particle counters used for detection are not specific to the tracer aerosols, and the resolution of detection is only in the µg/m³ range.

Another approach, by Pyankov et al.,¹⁵ uses influenza A virus as a tracer and a personal sampler for detection. However, they focus on introducing a new method of detection rather than a routine tracer entity for room characterization. Any approach using live viruses is not routinely feasible in a real-world setting as it requires high safety precautions, well-trained personnel, and extended experimental preparation time.

Further physical tracers,^{16–22} typically used outdoors and for environmental tracing, exist, but they, too, face challenges related to either environmental background concentrations, hazard, and toxicity or non-specificity of the detection method.

None of the reviewed tracing methods are suitable for complex multi-tracing or offer an extensive platform for flexible tuning of properties. There is still an unmet need for additional tracing methods allowing for direct and reliable experimental real-world characterization of indoor spaces regarding (bio-)aerosol distribution. Hence, novel tracing materials that are cost-effective, non-toxic, and can be detected with a high sensitivity and specificity in a broad range of scenarios would be a valuable addition to existing tools.

We propose a new approach using aerosolized silica particles with encapsulated DNA (SPED) to analyze aerosol dynamics. These submicron particles stand out in their ease of synthesis, durability, and quantitative analysis with a low detection limit, as described by

Practical Implications

For the first time, silica particles with encapsulated DNA were used to characterize a confined indoor space regarding position- and ventilation-dependent effects of aerosol distribution. The method described here introduces SPED as a novel, non-toxic, low-impact, cost-effective, and easy-to-use aerosol tracing platform that can be used to examine real-world environments. The mobile setup presented here as a proof of concept shows that SPED can be aerosolized and recaptured, followed by highly sensitive quantitative barcode-specific PCR analysis. The results revealed that this tracing method can detect position-dependent differences in exposure and ventilation effects influencing distribution dynamics. In the future, SPED could be engineered to exhibit custom-designed properties and be employed within a wide range of setups and high-capacity multi-tracing combinations.

Paunescu et al.²³ Silica is recognized as 'safe' by the FDA and used as food additive as well as in medicine.²⁴ A previous study shows no observable adverse effects of SPED in standard ecotoxicity assays²⁵ and exposure of rats to airborne silica nanoparticles in concentrations up to 86 mg/m³ over several days did not lead to any toxicological symptoms.²⁶ As previously shown, the employed sol-gel synthesis leads to a uniform size distribution with each particle carrying a comparable number of DNA sequences.^{23,27} The encapsulated artificial DNA, or 'DNA barcode', is not present in the environment and can be detected with an extraordinarily high selectivity and sensitivity using real-time quantitative polymerase chain reaction (qPCR). This powerful detection method is known for its use in diagnostics to identify various pathogens, for example in food,²⁸ water,²⁹ and also aerosols.^{15,30,31} Likewise, SPED are

TABLE 1 Literature overview listing aerosol tracing methods for room characterization

Method	Tracer type	Tracer size	Means of aerosol generation/ distribution	Means of collection	Method of quantification
Schade et al. ¹⁴	DEHS ^a (in conjunction with CO ₂)	300 nm (maximum of size distribution)	Aerosol generator with integrated pump	Handheld particle counters	Optical light scattering (particle counter)
Pyankov et al. ¹⁵	Influenza A virus	n/a	3-jet Collison nebulizer	Bioaerosol sampler	PCR
LUQAS study ⁷³	Artificial saliva	n/a	Particle generator ^c	No collection	LDA/PDA and PIV
Van Rijn et al. ⁷⁴	Glycerol/ethanol	Same size distribution as cough droplets	Specially designed spray nozzle	No collection	SprayScan [®] laser sheet
Knibbs et al.	CO ₂	No droplets	No aerosols	No collection	CA-10 CO ₂ analyzer
This work	Silica-encapsulated DNA	Ca. 150 nm ^d	Airbrush gun	BioSampler Impinger	PCR

^aDi-ethylhexyl-sebacate.

^bFor the devices used in the study, no lower limit of detection is available. For the Fidas[®] Frog, used for position-dependent counting, a measuring range of 0–20 000 000 p/L is indicated by the manufacturer. For the PCE devices used for time-resolved sampling, the resolution is 1 µg/m³ according to the product information, meaning the limit of detection would be equal to that, or higher.

^cNot further specified, but image provided in study.

^dMean of hydrodynamic size distribution of tracer agent in suspension.

already established as tracers for liquids and surfaces and have been used to characterize aquifers,³² assess pesticide drift,³³ track and trace commodities,^{34,35} and observe trophic interactions within food webs.³⁶ Moreover, they have been employed as surrogate tracers for bacteria in a hospital environment.³⁷ Similarly, bacteriophages, *C. difficile* spores, and cauliflower mosaic DNA have been employed for surface tracing in various settings, sometimes in conjunction or compared with fluorescent markers.^{38–43} The DNA sequences are unique identifiers, offering a great advantage over other methods by allowing for specific detection. In the case of SPED, the sequences are synthetic, not bound to an organism, can be changed at will and are more robust through silica encapsulation. Since even a sequence length of just 60–100 nucleotides offers trillions of potential barcodes, a system with nearly unlimited multi-tracing capabilities is conceivable. Thus, the advantages of highly sensitive PCR detection can be implemented without the need for tracing organisms.

In this study, SPED are dispersed as aerosols and recollected using commercial biosampler impingers, which offer a proven method to capture aerosols.^{44–46} The DNA is then released from its silica protection and quantitatively analyzed. The overall principle is summarized in Figure 1. Two batches of SPED, S1 and S2, were used, which differ in the sequence of their DNA barcode. Working with these unique identifiers has the advantage that they enable a wide variety of experiments, largely eliminating contamination between experiments and permitting scenarios such as simultaneous multi-source sampling.

2 | MATERIALS AND METHODS

2.1 | Synthesis of SPED

Silica particles with encapsulated DNA synthesis (S1 and S2) was performed in adaption of Paunescu et al.²³ and characterized regarding DNA load, size, and shape. For DNA encapsulation, 4 × 4 ml of silica

nanoparticles (110 nm, 50 mg/ml in isopropanol; Pinfire) per batch were surface-functionalized in 4 separate falcon tubes by adding 40 µg of N-trimethoxysilylpropyl-N,N,N-trimethylammonium chloride (TMAPS) (50% wt in methanol; abcr) followed by 12 h of stirring at 900 rotations per minute (rpm) at room temperature. For DNA adsorption to the surface, a 2 ml batch of 150 ng/µl corresponding annealed DNA (sequences see Table 2; Microsynth AG) was added to 200 ml ultrapure water (mQ; type 1, 18.2 MΩ·cm at 24°C, Milli-Q®; Merck). 0.4 g of the functionalized particles were added to the DNA solution and shaken for 10 s. Subsequently, 4 µl TMAPS were added, then the mix was shaken and sonicated for 20 s. Next, 62.5 µl of tetraethyl orthosilicate (TEOS) (≥99.0%; Sigma-Aldrich) were added, followed by 5 h of shaking at 600 rpm on a mixer (Vibramax 100; Heidolph) with a universal clamping attachment (VX 8; IKA). In a further step, 10 ml isopropanol and 5.9 ml TEOS were mixed with 484.1 ml mQ water and combined with the previous mixture. The batch was again stirred at 600 rpm for 4 days, before washing twice with water.

2.2 | SPED characterization

DNA load was determined by photospectrometric measurement of DNA concentration in solution before and after encapsulation (NanoDrop 2000c; Thermo Fisher Scientific). Particle size distributions of SPED in suspension (prepared in mQ water at ~2 mg/ml) were measured using an analytical photocentrifuge (dispersion analyzer LUMiSizer, light source 470 nm; LUM GmbH.). Transmission profiles were recorded for 2 h in time intervals of 5 s, at a rotational speed of 3000–4000 rpm. Statistical data analysis was performed by SEPView® software (LUM GmbH). Scanning electron microscopy (SEM) was performed on NovaNanoSEM450 device (Field Electron and Ion Company) with samples loaded on a TEM grid.

Time-/location-resolved results	Lower limit of detection	Specificity for tracer	Tracer occurring naturally in environment	Toxicity/ infectious-ness at working conc.	Multi-tracing possible
Yes/yes	1 µg/m ³ resolution in time-resolved measurement ^b	No	No (DEHS)/ yes (CO ₂)	No	No
No/no	n/a	Yes	Yes	Yes	No
Yes/yes	n/a	No	No	No	No
Yes/yes	n/a	No	No	No	No
Yes/no	n/a	Yes	Yes	No	No
Yes/yes	<0.1 ng/m ³	Yes	No	No	Yes

FIGURE 1 Schematic description of the tracing method using SPED. DNA is encapsulated in silica particles to form SPED, which are then aerosolized using an airbrush gun. The air is sampled using glass biosamplers. DNA is released from the collected SPED and quantified by qPCR

TABLE 2 DNA sequences encapsulated in silica for SPED synthesis

Name	Sequence
S1 forward	TATGCGCCTTTTACTCTTATAGGTATCCTGTTGCTGGCACTTTTTCTAGCAAAGTCTTCTCCT
S1 reverse	AGGAGAAGACTTTGCTAGAAAAAAGTGCCAGCAACAGGATACCTATAAGAGTATAAAGGCCGATA
S2 forward	TAGCTCGTTCATAGAATCACTTCGCCGTACTCAACGTAGTGTTTTGTTTAGCTCAAACAGGTT
S2 reverse	AACCTGTTTGAGCTAAACAAAAACCACTACGTTGAGTACGGCGAAGTGATTCTATGAACGAGCTA

2.3 | Aerosolization and droplet characterization

The aerosol experiments were performed in a 277 m³ laboratory running at a ventilation throughput of 1840 m³/h. A technical layout of the room can be found in the Appendix S1, Figure S2d. SPED-containing aerosols were generated by a commercial airbrush gun with a 0.35 mm nozzle (type AFC-101A; Conrad Electronic) using an air pressure of 1 bar, mounted to a height of 1.6 m on a mobile stand. Phase Doppler anemometry was used to characterize droplet size distribution as created by the airbrush gun used for aerosol generation. PDA was measured on a setup using a Phase Doppler anemometer (TSI Inc.) operated with a water-cooled 5 W argon-ion laser (Innova 70, LA-70-5), generating a green and blue laser beam with a wavelength of 514.4 and 488 nm, respectively. To measure droplet size distribution, pure water was used, at a flow rate of ca. 2 ml/min. The point of measurement was 2 cm frontal to the nozzle.

For each sampling, 5 ml of a 0.2 mg/ml aqueous suspension of the respective SPED species were nebulized over the course of approx. 1 min. For the treatment and wash protocol of SPED before dispersion, refer to the Appendix S1, Section S1.4.

2.4 | Aerosol capture

For aerosol capture, biosamplers, also called flow impingers (Aquaria SRL and SKC Ltd), were filled with 20 ml of ultrapure water and equally mounted to a height of 1.6 m. For photographs and descriptions of the different components refer to the Appendix S1, Figures S3 and S4. Before aerosol dispersion, zero sampling was performed, which consisted of sampling the empty room using the same setup and conditions.

Air samplers were used in combination with 40 l/min piston vacuum pumps (Ningbo Nuolin Mechatronics Co Ltd), whereby the airflow rate was reduced to 12.5 l/min using needle valves. If not indicated otherwise, sampling time was 2 h post-dispersion. For time-resolved experiments, the pumps were briefly stopped at given time points, each time removing 200 µl of the respective capturing liquid from the samples for analysis.

2.5 | Quantitative PCR

Immediately after sampling, the collection liquid from each flow impinger was transferred into a 50 ml falcon tube and sonicated for 10 min in an ultrasound bath (model Sonorex Digitec DT 31H; Bandelin) at room temperature. A total of 200 µl were used for further processing and treated with 4 µl buffered oxide etch, consisting of 0.03 wt% ammonium hydrogen difluoride (NH₄FHF, pure; Merck) and 0.02 wt% ammonium fluoride (NH₄F, puriss.; Sigma-Aldrich). After adding the etching reagent, the mix was vortexed and sonicated again. For the subsequent polymerase chain reaction, 5 µl of the broken-down SPED solution were mixed with 10 µl of KAPA SYBR FAST qPCR master mix universal (2x; Kapa Biosystems), 1 µl of each primer solution, and 3 µl of PCR-grade water (type 1, 18.2 MΩ·cm at 24°C, Milli-Q®; Merck). Primer sequences (see Table 3) were ordered in dry state from Microsynth AG and dissolved in PCR-grade water to a working concentration of 10 µM. Samples were run in technical triplicates on a LightCycler® 96 instrument (Roche Molecular Systems). The qPCR program consisted of a pre-incubation step (240 s, 95°C) followed by 40 cycles of 3-step amplification (2 s, 95°C; 12 s, 60°C; 4 s, 72°C). Cycle values were assigned automatically from qPCR fluorescence curves by the LightCycler®

96 SW 1.1 analysis program (Roche Molecular Systems) and converted to concentrations based on previously measured standard curves. For a more detailed description of the calculations and determination of experimental detection limits refer to Appendix S1, Sections S1.4–5 (Table 3).

2.6 | Fog experiment

For visualization of air and ventilation dynamics, fog was generated by a commercial fog machine (Model Rage 600I; BeamZ), used in conjunction with the liquid provided with the device (BeamZ fog fluid 1 L High-Density). The fog was dispersed according to instructions by the manufacturer from the same point of origin as the SPED. Per experiment, one load was dispersed, corresponding to 20–30 s of continuous operation. Qualitative analysis was performed by eye and documented in images and on video S1.

3 | RESULTS AND DISCUSSION

Characterization of the two SPED species (S1 and S2) used for the following tracing studies revealed a DNA load of 23 μg and 26 μg dsDNA/mg for S1 and S2 particles, respectively. The median hydrodynamic size was determined to be 133.5 nm for S1 and 160.6 nm for S2, as shown in Figure 2B. This is further confirmed by electron microscopy (Figure 2A), which shows the spherical shape and homogeneous size distribution of the particles. The selected core particle

size of 110 nm is similar to viruses like adenovirus,⁴⁷ influenza A,⁴⁸ or SARS-CoV-2,⁴⁹ all ranging between 90 and 120 nm in size.

In addition to the hydrodynamic diameter of SPED in suspension, the droplet size generated by the airbrush gun was determined using phase Doppler anemometry (PDA), the results of which are shown in Figure 2C. The measured particle count distribution yielded a droplet size of $10.8 \pm 4.4 \mu\text{m}$. There is abundant literature characterizing size distributions of droplets as created by the human respiratory tract.^{50–62} The reported average sizes range from submicrometer diameters up to several hundred micrometers. The size differs between types of respiratory action, i.e. regular breath, speech, coughing, or sneezing, and also greatly varies depending on study design variables, such as point of measurement, means of collection and detection, and statistical methods. In a study employing an optical detection method comparable to PDA to measure respiratory droplets in immediate proximity to the source, a median droplet size of 12.3 μm was found for coughing,⁶¹ which is a similar size range as generated by the airbrush gun.

Further, it is important to note that the droplet size distribution and the particle size distribution are two separate entities. The droplet size distribution is the factor determining which fraction of the dispersed particles become airborne after the carrier liquid evaporates, as described in a model developed by Wells et al.⁶³ This so-called Wells falling evaporation curve was later refined by Xie et al.,⁶⁴ showing that water-based droplets below a size of 60 μm evaporate quickly under standard indoor conditions and can remain suspended in the air for longer periods of time. Considering this, the entities collected are not primarily the droplets as dispersed but the aerosolized droplet nuclei. This is schematically shown in Figure 3. This phenomenon, too, is multifactorial and determined by interdependent parameters such as temperature, humidity, and the concentration of particulate matter suspended in the air. Generating an experimental setup that mimics human respiration is therefore not trivial and was not a primary focus of this study. Nevertheless, the droplet size distribution created by the airbrush gun falls within the previously reported range for respiratory droplets and SPED have a similar hydrodynamic volume as a typical virus as base properties

TABLE 3 Primer sequences used for qPCR

Primer name	Sequence
S1 forward	ATGCGCCTTTATACTCTTA
S1 reverse	GGAGAAGACTTTGCTAGAA
S2 forward	AGCTCGTTCATAGAATCAC
S2 reverse	ACCTGTTTGAGCTAAACAA

FIGURE 2 (A) Scanning electron microscopy (SEM) image of SPED species S2. Scale bar is 500 nm. (B) Hydrodynamic size distribution of two batches of SPED (S1 and S2) by analytical photocentrifugation. For S1, the median is at 133.5 nm, for S2 at 160.6 nm. (C) Diameter of droplets generated by the airbrush gun, as measured by phase Doppler anemometry. The arithmetic mean droplet size is $10.8 \pm 4.4 \mu\text{m}$

for the following measurements. Figure 4A shows a summary of the setting used to investigate and quantify the relative distribution of SPED aerosols indoors. 5 ml of a 0.2 mg/ml particle suspension were dispersed over the time course of ca. 1 min in a 277 m³ laboratory, running at a ventilation output of 1840 m³/h. This corresponds to roughly 6.6 air changes per hour (ACH) or roughly one air changes every 10 min. For aerosol capture, up to 5 flow impingers were mounted with the inlets at a height of 1.6 m. This height was chosen because it is similar to the level of the nose of an average worker in the laboratory. A graphic depicting the location of the samplers within the test space is shown in Figure 4B. For aerosol capture, each flow impinger was connected to a pump aspirating 12.5 l/min of ambient air, which is close to a human's minute ventilation for a low-level activity, such as driving a car.⁶⁵ This enables sampling of a significant amount of air over time, while still detecting local exposure rather than sampling the room globally. Through the focused airflow, aerosols are trapped in the collection water, which can directly be used for further processing and subsequent PCR-analysis. The setup is mobile and only requires electricity from regular sockets and a small-compressed air source for the airbrush gun. A battery-powered version of the same setup would also be feasible.

As a pilot test, a time-resolved experiment was conducted. For this experiment, two flow impingers were used, mounted 2 m and 6 m frontal to the origin of aerosol flow, respectively, and the test laboratory ran at regular ventilation output. 1 mg of SPED 1 was dispersed, and during a continuous sampling time of 120 min, 200 µl samples were removed from the initial 20 ml solution at 14 time points and analyzed separately for their SPED concentration. Figure 5A shows the integrated aerosol levels measured in the two

impingers over time. The first measurement, at $t = 0$, was taken before dispersion and is thus the experimental negative control, marking the minimal limit of detection (MLD) of $4.5 \cdot 10^{-9}$ mg/ml in the sampling solution for this experiment. Assuming a conservative physical collection efficiency for the impingers of 10%, as estimated from literature data,^{66–68} this corresponds to a particle concentration detection limit of $6 \cdot 10^{-2}$ ng/m³ of sampled air. Comparing this value to commercial monitoring systems for inorganic particulate matter 2.5 (PM_{2.5})^{69,70} the MLD of the present method is at least one order of magnitude lower. More information on the impinger efficiency, concentration measurements, and how detection limits were determined can be found in the Appendix S1, Section S1.5.

Following the aerosol generation, the amount of SPED collected rises sharply in both impingers, before reaching a plateau load in the range of $3 \cdot 10^{-4}$ mg and $2 \cdot 10^{-4}$ mg, respectively. At position 1, the maximum is reached within the first measurement increment after 3 min, whereas at a distance of 6 m, there is a distinguishable saturation curve reaching the plateau after an estimated 10–20 min, attributing to the fact that the aerosols reach the second sampling point with a delay due to further traveling distance. This position-dependent effect is also reflected in the total collection loads, which correspond to estimated position-specific dosages of 0.35% and 0.2% of the total amount of released SPED. Throughout the measurement time, there are some fluctuations; however, these deviations are within the expected range for PCR reactions, as discussed further below.

To further investigate position dependency, the total load over 120 min was measured under otherwise equal sampling conditions. In addition to the two impingers used in the previous experiments,

FIGURE 3 Schematic representation of exposure scenarios with evaporating droplets, comparing (A) virus-carrying aerosols as produced by a human cough with (B) the laboratory setup used in this study

FIGURE 4 (A) Schematic representation of room characterization experiments. SPED are dispersed in an indoor environment, recollected using one or several sampling devices and analyzed using quantitative PCR. (B) Room layout of the laboratory used for the experiments, schematically showing benchtops, sinks, fume hoods, and hallways. Distances of the sampling device locations from the source of dispersion are indicated

two locations with a 2 m offset relative to the airbrush gun's line of flight were sampled simultaneously (labelled positions 3 and 4 in Figure 4B). The results of three separate runs—two with barcode S1 and one with S2—are displayed in Figure 5B. To relate total load to the distance from the source, the data are always normalized to the 2 m frontal position, which is standing within the aerosol flow and therefore considered an experimental positive control. The data show that relative exposure is much lower at all three locations that are either further apart from the source or not in the direct line of flight. Additionally, the three experiments qualitatively compare with each other, independent of the DNA barcode used.

In a next step, we wanted to determine the influence of ventilation to aerosol distribution and clearance. We therefore repeated the previous measurements but turned off the ventilation system with doors and windows remaining closed, as before. Figure 5C compares total load after 2 h of sampling at the four previous sampling points with ventilation on and off, respectively, again relative to the position close to the point of aerosolization. This experiment revealed a more pronounced concentration difference between the individual sampling points. This result follows intuition, as in the non-ventilated scenario the airflow in the room is more stagnant as opposed to the forced convection found in the experiments under room ventilation.

To assess whether SPED could also be detected in an adjacent room, we placed an additional sampler in the atrium next to the test laboratory (position 5 from Figure 4B), separated by a closed door,

with a cubature of 30 m³. These results are displayed separately in Figure 5D, comparing position 5 to position 1. The expectation was that exposure in the atrium would only be minimal. Interestingly, the exposure is considerable in the ventilated scenario but very low without ventilation. Based on this outcome, we hypothesized that ventilation dynamics could lead to a suction effect from the main room to the atrium. The layout of the ventilation piping (see Appendix S1, Figure S2d), which shows a connection between the two rooms, is compatible with this theory.

As these results were unexpected, we wanted to reconfirm them using another method. We therefore used a fog machine to visualize air dynamics and to add qualitative evidence to the previous results. The fog was generated at the same position as the previous source, and a video camera was installed in the atrium. Indeed, with ventilation running, fog could be visually detected with a delay of ca. 5 min (see Video S1) seemingly coming from the ventilation pipe. The fog does not allow for quantification or localized detection and has a comparatively high detection limit, but it was sufficient to qualitatively confirm our measurements. This shows that the method introduced here is suitable to detect unexpected effects in the distribution dynamics of aerosols in a real-world setting, as caused by a specific ventilation system. There are indications that ventilation systems and air-conditioning can contribute to the spreading of airborne diseases instead of preventing it,¹⁰ which is why it is already recommended to replace

FIGURE 5 Time-, position-, and ventilation-dependent effects as measured by air sampling after dispersion of SPED aerosols. Error bars are calculated from standard deviations of PCR triplicates. Position numbers refer to sampler location in the room as depicted in Figure 4b. (A) Time-dependent measurement using two air samplers. The y-axis indicates the integrated amount of collected SPED (in mg) over a total time of 120 min. The measurement point at 0 min was taken immediately before dispersion as a zero control. The first five measurements are zoomed in for clarity. (B) Measurement of position-dependent relative exposure of 4 impingers. Two experiments were performed with particles S1 and a third using particles S2. The values reflect the total amount of particles collected after a 2-h sampling time. The data are normalized to the first position in each set. (C) Total load relative to position 1 per impinger with (data from S1 experiment 2) and without ventilation under equal conditions as in (C). (D) Same as (C) but comparing position 1 to an additional position in the adjacent atrium (position 5), which is additionally zoomed in for clarity

air recirculation by increased inflow of outdoor air.⁷¹ The results show that the described method is able to measure such effects and could, for example, be employed in finding appropriate ventilation settings to limit aerosol spread.

The present proof of concept shows that a simple setup can be used to reliably measure time-, position-, and ventilation-dependent relative aerosol loads indoors using different DNA barcodes. Current limitations of the sampling and detection method are a direct result of PCR analysis, which requires normalization to achieve quantitative results. Furthermore, PCR data are logarithmic to the concentration levels. Consequently, result quantification requires a range of control experiments, and small concentration differences are more difficult to detect, requiring numerous sample replicas. Furthermore, the individual measurements were conducted in a real-world environment under similar conditions but without perfect control of external parameters. Factors such as relative humidity, temperature, and the presence and concentration of other particulate matter in the ambient air can influence aerosol dynamics and lead to a difference in absolute aerosol levels accessible to measurement.

Despite the limitations mentioned above, the SPED-based method introduced here has several important benefits, the

combination of which makes it novel and unique among aerosol tracers. Important features are the versatility, sensitivity, and specificity toward the sampled barcode and suitability to detect effects related to air circulation. The silica layer protects the DNA from physical and chemical damage,²³ which is an additional advantage over other DNA-based tracers and makes SPED suitable for indoor and outdoor use alike. The costs of SPED are estimated at 500 USD/g, which benchmarks the price of particles for a single experiment of the scale discussed here at less than 1 USD.

Even though flow impingers were used in this experimental setup, the use of SPED in combination with other means of aerosol capturing is conceivable. This includes surface sampling, even though in past studies³⁷ high local SPED concentrations were required for successful swabbing. Furthermore, while the airbrush gun used in this pilot study presents a cost-effective, user-friendly solution, SPED could equally be used in combination with advanced dispersion systems to, for example, more specifically mimic human breathing or coughing.⁷² It has already been shown that SPED detection is possible at a single particle level in solution.²⁷ Thus, when limiting barcodes to single use and with a further focus on optimization of the respective sampling and measurement conditions, an additional improvement of the detection

limit in the range one or two orders of magnitude is conceivable. Consequently, the flexible barcodes offer countless combination possibilities in the future measurements, such as simultaneous surface and air sampling with multiple sources and detectors at extremely low detection limits.

4 | CONCLUSIONS

In conclusion, this study presents a simple, cost-effective setup for investigating aerosol distribution using a novel tracing agent. The SPED-based platform offers a foundation for a range of real-world tracing scenarios at low detection limits using qPCR as method of analysis. Future applications are the study of aerosol flow in complicated architectural settings and the dynamics of free convection resulting from heating, daytime effects, and door/window arrangements. In the near future, SPED could additionally be tuned in size, made biodegradable, or be designed to mimic specific pathogens or hazardous pollutants. These are ideal prerequisites for developing a robust platform to examine places of interest regarding potential health hazards and environmental risks.

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CONFLICT OF INTEREST

The authors declare the following competing financial interest(s): R.N.G. and W.J.S. declare a financial interest in the form of IP on DNA encapsulation licensed to Haelixa AG, of which R.N.G. and W.J.S. are shareholders. A.M.L. and J.K. have no competing financial interests.

PATIENT CONSENT STATEMENT

No patients were involved in this study or any experiments part of it.

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PEER REVIEW

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SUPPORTING INFORMATION

Additional supporting information may be found in the online version of the article at the publisher's website.

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